RESEARCH ARTICLE



Temperature effects on root exudation in mature beech (*Fagus sylvatica* L.) forests along an elevational gradient

Christoph Leuschner[®] · Timo Tückmantel · Ina C. Meier

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Abstract

Aims Root exudation may have a large impact on soil biological activity and nutrient cycling. Recent advances in *in situ*-measurement techniques have enabled deeper insights into the impact of tree root exudation on rhizosphere processes, but the abiotic and biotic controls of exudation rate remain poorly understood. We explored the temperature dependence of root exudation in mature beech (*Fagus sylvatica* L.) trees.

Methods We measured fine root exudation in seven beech forests along an elevational gradient (310– 800 m a.s.l.) and related carbon (C)-flux rates to mean daily temperature, actual precipitation, mean summer temperature (MST) and precipitation (MAP), soil moisture (SWC), and stand structure.

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C. Leuschner (⊠) · T. Tückmantel · I. C. Meier Plant Ecology, University of Goettingen, Untere Karspüle 2, 37073 Göttingen, Germany e-mail: cleusch@gwdg.de

Present Address: I. C. Meier Functional Forest Ecology, Universität Hamburg, Haidkrugsweg 1, 22885 Barsbüttel, Germany Results Average mass-specific exudation (averaged over all sampling dates) ranged from 12.2 µg C $g^{-1} h^{-1}$ to 21.6 µg C $g^{-1} h^{-1}$ with lowest rates measured at highest elevations and peak rates at mid-elevation (490 m). Regression analyses showed a highly significant positive effect of site-specific daily air and soil temperature on exudation rates (p < 0.01) with an average increase by 2 µg C g⁻¹ h⁻¹ per 1 °C-temperature increase, while the relation to mean summer or annual temperature and mean temperature of the measuring year was less tight. Exudation decreased with increases in mean annual precipitation and soil moisture, but increased with increasing stem density. Conclusions The root exudation rate of beech trees roughly triples between 10 °C and 20 °C mean daily temperature, evidencing a large temperature influence on root-borne C flux to the soil.

Keywords Air temperature · European beech · Fine root morphology · Mass-specific exudation · Precipitation · Soil moisture · Soil temperature

Introduction

Climate warming affects biogeochemical processes and the carbon (C) cycle of forests through various direct and indirect interactions, notably temperature effects on photosynthesis, plant respiration, soil organic matter (SOM) decomposition, nutrient diffusion in soil, and root nutrient uptake kinetics, and the abundance and activity of mycorrhizal partners (Norby et al. 2007; Yin et al. 2013b). While much experimental research with trees has addressed warming effects on photosynthesis, respiration, belowground C allocation and fine root dynamics (Bai et al. 2010; Graham et al. 2014; Liang et al. 2013; Majdi and Ohrvik 2004), and soil microbial activity and SOM decomposition (von Lützow and Kögel-Knabner 2009), much less is known about the temperature dependence of rhizodeposition, i.e. the loss of carbon from roots (Yin et al. 2013b). Rhizodeposition, which includes exudation (the leakage of soluble organic compounds from living root cells), the secretion of mucilage and other organic substances, and the shedding of dead cells and tissues from roots, represents a net flow of C from roots to the soil which can have profound effects on C and nutrient cycling in forest ecosystems (Jones et al. 2009). Root exudation is thought to be a key mediator in plant-soil interactions that influences the composition and activity of soil microbial communities and thus soil enzyme activity, SOM decomposition and nutrient cycling (Cheng et al. 2014; Fransson and Johansson 2010; Gougherty et al. 2018; Jones et al. 2004; Nguyen 2003; Pausch and Kuzyakov 2017). Estimates of the size of the C flux with root exudation in forest ecosystems vary widely from a few percent up to 21% of net primary productivity (Badri and Vivanco 2009; Jones et al. 2004; Kannenberg and Phillips 2017; Meier et al. 2020; Pinton et al. 2007), or up to a third of the photosynthetic carbon gain of tree saplings (Liese et al. 2018). It is not well understood how an increase in temperature influences exudation rate, as this C flux is dependent on a multitude of biotic and abiotic factors and is rarely quantified under field conditions (Yin et al. 2013b). In fact, it may represent the most uncertain part of the C cycle in ecosystems, especially in mature forests (Price et al. 2012).

Biotic factors influencing root exudation are carbohydrate availability and allocation to roots, plant phenology and development stage, root morphology (notably specific root length, SRL), the type of mycorrhizal fungi, and the presence of pathogenic soil microbes (Kuzyakov 2002; Meier et al. 2013; Neumann and Römheld 2007; Tückmantel et al. 2017) as well as root membrane conductivity. Abiotic factors that have been found to influence root exudation rate include radiation through its influence on photosynthesis (Nakayama and Tateno 2018), deficiency of nutrient elements such as phosphorus (P) and nitrogen (N) (Meier et al. 2020; Yin et al. 2014), soil moisture (Brunn et al. 2022; Jakoby et al. 2020; Preece et al. 2018), and temperature (Yin et al. 2013a). The majority of root exudates are believed to be passively lost from the root, thus being proportional to the root – soil solute concentration gradient (Jones et al. 2004). This suggests a possible influence of microorganism absorption on exudation rate (Canarini et al. 2019).

The few studies that have addressed temperature effects on tree root exudation employed either experimental soil warming of seedling cultures (Karst et al. 2017; Uselman et al. 2000; Yin et al. 2013a, b) or compared trees of the same species at sites differing in mean temperature (Yang et al. 2020). While most studies with trees indicate higher exudation under elevated temperatures, higher exudation at lower temperatures was also reported (Karst et al. 2017; Yang et al. 2020). Due to the multitude of possible influential factors and the difficulties associated with measuring tree root exudation under field conditions, our understanding of climate warming effects on exudation is very limited (Wang et al. 2021). This weakens our capacity to predict changes in soil biological activity and C and nutrient cycling in forest soils in a future warmer world.

In this study, we investigated changes in root exudation in seven mature forests of European beech (Fagus sylvatica L.) along an elevation transect (310 - 800 m a.s.l.) on acidic soil with the aim to explore the responses of root exudation of this tree species to both prevailing thermal conditions and varying mean annual temperature (8.4 - 6.0 °C) at the site. Earlier research in Central European beech forests with an insitu cuvette-based method showed that mass-specific exudation rates decrease greatly from the topsoil to the subsoil (Tückmantel et al. 2017) and increase with soil acidity and N deficiency of the soil (Meier et al. 2020). To capture the assumed temperature effects, we here extend the sample of studied beech forests to an elevational transect on base-poor bedrock from the colline to the montane zone with a marked temperature decrease and precipitation increase, while soil pH (pH(CaCl₂ in 0-10 cm mineral soil: 3.1 - 3.7) and soil N availability $(15.2 - 19.9 \text{ g s}^{-1})$ varied only little. We hypothesized that the root exudation rate increases with increasing mean daily temperature due to an overall increase in the tree's metabolic activity,

while the cumulative annual C flux with exudation is primarily determined by growing season length, which increases from the colder to the warmer stands.

Material and methods

Study sites, climate, and geology

The study was conducted in seven European beech forests of mature age along an elevational gradient between 310 and 800 m a.s.l. on the eastern slopes of the Rothaar Mountains in the state of Hesse, central Germany. The study region has a cool-temperate humid climate with MAT decreasing from 8.4 to 6.0 °C and mean annual precipitation (MAP) increasing from 600 to 1200 mm yr⁻¹ from the colline to the montane zone due to orographic lift of air masses that mostly arrive from western directions (German Weather Service, period 1981–2010; Table 1).The elevation transect had a length of approx. 30 km in east–west direction and ranged from the colline/

submontane to the montane belt, covering a gradient in mean annual temperature (MAT) of about 2.4 °C and a MAP gradient of 600 mm yr^{-1} . The studied beech forests are thus exposed in downslope direction to increasingly warmer and drier summers, as is predicted to happen in the course of climate warming in the twenty-first century in Central Europe and elsewhere (Kaspar et al. 2017). By employing a space-for-time substitution, this setting may allow rough estimates of anticipated future changes in tree root exudation with climate warming. Besides temperature, we measured soil moisture and several soil chemical parameters as well as fine root biomass (FRB) and fine root morphological traits in order to relate exudation to possible controlling abiotic and biotic factors. All stands belonged to the Luzulo-Fagetum forest community (beech forests on acidic soils) and stocked on acidic bedrock (Triassic sandstone or Paleozoic clay shale) in level to slightly inclined terrain (Table 1). In the forests, study plots of 30 m \times 30 m size were selected in sections with closed canopy. While mean diameter at breast height

Table 1 Location and physiographic characteristics of the seven study sites along the elevational gradient in central Germany

Elevation	m a.s.l	310	380	490	560	600	690	800
Longitude	[E]	08° 55'	08° 56'	08° 48'	08° 45'	08° 42'	08° 37'	08° 33'
Latitude	[N]	51° 19'	51° 18'	51° 17'	51° 18'	51° 17'	51° 16'	51° 17'
Inclination/ Exposition		10° SE-NW	5° NE-E	10° NW–SE	15° NW-NE	20° NE-E	20° N-Е	5° SE
Vegetation type		LF	LF	mon. LF	mon. LF	mon. LF	mon. LF	mon. LF
Tree age	[yr]	129–149	81-107	146-156	158-180	98-133	169–189	162–192
Stem density	$(n ha^{-1})$	311	578	267	489	267	150	250
Mean tree height	(m)	32.7	23.5	13.2	24.0	24.9	21.8	20.0
Mean dbh	(cm)	44.8	31.5	32.6	34.9	35.4	44.6	41.6
Growing season length	days $\geq 10 \ ^{\circ}\text{C}$	171	170	159	142	151	130	125
MAT	[°C]	8.1	8.4	7.9	7.4	7.7	6.6	6.0
Mean temperature 2014	[°C]	9.6	9.9	9.4	8.9	9.1	8.1	7.5
Mean temperature. 2015	[°C]	9.1	9.3	8.9	8.3	8.7	7.6	6.9
MST	[°C]	14.6	14.5	14.1	14.8	14.2	13.8	13.3
MAP	[mm]	605	643	691	887	951	1155	1209
Annual precipitation 2014	[mm]	706	707	764	845	839	1062	1115
Annual precipitation 2015	[mm]	571	561	621	673	696	974	1054

Vegetation type: LF—Luzulo-Fagetum; mon. LF—montane Luzulo-Fagetum. Tree age was retrieved from forest inventory data (Waldeckische Domanialverwaltung). Stem density, mean tree height, and mean diameter at breast height (dbh) were determined in August 2014. All climate data were derived from DWD (Deutscher Wetterdienst, German Weather Service, Offenbach, Germany), interpolated between neighboring measuring stations, and corrected for altitude. Climate data refer to multi-annual means (1981–2010) or annual means (temperature) or sums (precipitation) of the study year (for 2014 and 2015). MAT – mean annual temperature; MST – mean annual summer temperatures (May-Sept.) in the 1981–2010 period; MAP – mean annual precipitation in the 1981–2010 period

(DBH) varied only between 32 and 45 cm in the seven stands and tree ages ranged mostly between 100 and 180 years according to information from the forest offices, stem density was more variable (150-578 stems ha⁻¹). Mean tree height as measured in 15 trees per plot with a Vertex III height meter (Haglöf, Längsele, Sweden) with at least three measurements taken per tree from different directions decreased with elevation from 33 to 20 m.

During the study period from spring 2014 to winter 2015, air and soil temperature were continuously measured in 2015 with iButton sensors (Maxim, Dallas, USA) installed at 1.5 m height above the forest floor and in the topsoil (3 cm depth) in the seven stands. The sensors were read every 60 min. Dendrometer tapes (type D1, UMS, Munich, Germany; precision of 0.1 mm) were permanently installed at 1.5 m height on 15 trees per plot to determine DBH and annual stem diameter increment through annual DBH recording.

Soil chemical and physical analyses

To characterize soil chemical factors, each five samples were collected in summer 2015 from the uppermost 15 cm of the soil with a 6.6 cm-diameter corer at random position in the study plots. Subsequently, the thickness of the organic layer was measured in the cored hole. All soil samples were separated into organic layer and mineral topsoil material (0-10 cm), transferred to the laboratory in a cooling box, sieved (<5 mm for organic layer material; <2 mm for mineral soil), and stored in polyethylene bags at 4 °C for further processing. Subsamples were analyzed in field-moist condition for pH (measured in H₂O: 10 g fresh soil in 25 ml deionized water, or in CaCl₂: 10 g soil suspended in 0.01 M CaCl₂) after 1 h of equilibration. Additional subsamples were dried (60 °C, 48 h), ground, and analyzed for total carbon and nitrogen concentrations through gas chromatography with an elemental analyzer (vario EL III, Elementar Analysensysteme GmbH, Hanau, Germany). Since all soils were highly acidic, total C content was assumed to be organic C (SOC). The total P content was determined by ICP-OES analysis (Perkin Elmer Optima 5300 DV) after acid-pressure digestion (65% HNO₃ at 195 °C for 6 h) of the ground soil samples. Plant-available phosphorus was estimated with the resin bag method according to Bowman and Cole (1978) using Dowex $1 \times 8-50$ anion exchange gel (Dow Water & Process Solutions, USA) that was placed for 16 h in a solution of 1 g of soil material suspended in 30 ml water (Sibbesen 1977). Extracted P was re-exchanged with NaCl and NaOH solutions and the P concentration measured in a spectrophotometer (Libra S 21, Biochrom, UK) at 712 nm after adding 5 mM hexaammonium-heptamolybdate solution (Murphy and Riley 1962).

The water content of the topsoil was determined gravimetrically in each five soil samples collected at random position in the plots every month from March to December 2015. The sampling in August and September was conducted synchronously with the collection of root exudates.

Root exudate collection

Root exudates were collected in three sampling campaigns in July of 2014, August 2015, and September 2015 in nine soil pits excavated at each site in at least 3 m distance to the nearest mature beech tree, employing the cuvette-based in situ-collection approach (after Phillips et al. 2008; Freschet et al. 2021). Beech fine root strands were carefully extracted from the uppermost 10 cm of the soil profile and cleaned with fine forceps and deionized water. Since organic layer thickness increases with elevation (Table 2), the sampled roots were at the uppermost sites 6 and 7 entirely located in the thick organic OF and OH layers, while a greater part of the studied root strands grew in the uppermost humus-rich mineral soil Ah-horizon in the lower plots (1–5) with somewhat thinner organic layers. After cleansing, roots were placed overnight in moist, sandy soil to allow recovery from the excavation process. On the next day, the living terminal root systems (average cumulative length of all parts of the strand c. 16.5 cm and mean diameter c. 0.45 mm) were placed into root cuvettes filled with sterile 2 mm-diameter glass beads to simulate the porosity of the soil and mechanical impedance in a matrix free of carbon. The beads covering the root were moistened with a sterile carbon-free dilute nutrient solution (0.5 mM NH₄NO₃, 0.1 mM KH₂PO₄, 0.2 mM K₂SO₄, 0.15 mM MgSO₄, 0.3 mM CaCl₂) used as a culture medium.

The roots in the cuvettes were allowed to recover for 48 h before flushing and cleaning the cuvettes 3 to 5 times with culture medium using gentle, lowpressure vacuum (\leq -0.2 bar) induced by a syringe. New sterile culture medium (c. 40 mL) was added **Table 2** Soil physical and chemical characteristics of the organic layer and the mineral topsoil (0-10 cm) of the seven beech forests along the elevational gradient (means and SE). Results for soil water content (SWC) are the means of gravimetric determination on nine sampling dates with n=5 sam-

ples per site. Different upper-case and lower-case letters indicate significant differences between sites at p < 0.05 (*p*-values adjusted by the Benjamini–Hochberg procedure for multiple comparisons)

Elevation	m a.s.l	310	380	490	560	600	690	800
Parent material		Sandstone	Sandstone	Clay shale	Clay shale	Clay shale	Clay shale	Clay shale
Geological epoch		1 BU	1 BU	1 K	u D	u D	m D	m D
Org. layer thickness	cm	4.5 ± 0.26	5.2 ± 0.35	7.7 ± 0.42	7.8 ± 0.39	7.9 ± 0.66	10.9 ± 0.65	10.5 ± 0.49
Org. layer mass	[kg m ⁻²]	7.97 ± 1.05	7.56 ± 1.18	8.41 ± 0.86	11.05 ± 1.01	11.21 ± 1.76	13.90 ± 1.38	15.03 ± 2.44
Texture								
0–10 cm		Sandy loam	Sandy loam	Sandy loam	Loamy silt	Loamy silt	Silt loam	Sandy Loam
Bulk density								
Org. layer	[g cm ⁻³]	0.18 ± 0.02	0.13 ± 0.02	0.12 ± 0.01	0.15 ± 0.01	0.14 ± 0.02	0.12 ± 0.01	0.14 ± 0.02
0–10 cm	[g cm ⁻³]	1.32 ± 0.02	1.36 ± 0.04	1.11 ± 0.03	1.07 ± 0.07	0.99 ± 0.11	0.75 ± 0.06	0.76 ± 0.04
SWC								
Org. layer	[wt %]	46.9 ± 2.1	41.4 ± 1.9	59.1 ± 1.6	57.8 ± 1.6	52.4 ± 1.7	61.3 ± 1.2	66.9 ± 1.1
0–10 cm	[wt %]	22.8 ± 0.9	18.7 ± 0.8	26.1 ± 1.3	27.1 ± 1.3	29.4 ± 1.1	38.1 ± 0.7	43.6 ± 1.4
pH (H ₂ O)								
Org. layer		$4.3^{A} \pm 0.03$	$4.7^{A} \pm 0.38$	$4.2^{AB}\pm0.02$	$3.9^{\rm B}\pm0.07$	$4.2^{\rm AB}\pm0.03$	$4.2^{\rm AB}\pm0.07$	$4.1^{\mathrm{AB}} \pm 0.04$
0–10 cm		$3.7^{b} \pm 0.02$	$4.6^{a} \pm 0.09$	$3.8^{a} \pm 0.06$	$3.8^{a} \pm 0.12$	$3.7^{b} \pm 0.09$	$3.7^{b} \pm 0.03$	$3.7^{b} \pm 0.18$
pH (CaCl ₂)								
Org. layer		$3.7^{A} \pm 0.05$	$4.3^{A} \pm 0.34$	$3.6^{A} \pm 0.05$	$3.3^{\rm B}\pm0.09$	$3.4^{A} \pm 0.06$	$3.3^{B} \pm 0.04$	$3.3^{B} \pm 0.13$
0–10 cm		$3.4^{ab}\pm0.04$	$3.5^{ab}\pm0.05$	$3.5^{a} \pm 0.03$	$3.3^{ab}\pm0.05$	$3.4^{ab}\pm0.03$	$3.3^{ab}\pm0.02$	$3.1^{b} \pm 0.18$
C _{tot}								
Org. layer	[mg g ⁻¹]	$143.7^{D}(\pm 9.6)$	$175.7^{\rm CD}\pm8.4$	$255.4^{AB} \pm 11.6$	253.6 ^{AB} 12.1	$215.0^{\rm BC} \pm 10.8$	$276.5^{A} \pm 14.2$	$304.4^{A} \pm 10.5$
0–10 cm	[mg g ⁻¹]	$26.1^{\circ}(\pm 1.7)$	$28.1^{\circ} \pm 1.4$	$52.9^{b} \pm 7.2$	$63.3^{b} \pm 7.8$	$56.9^{b} \pm 3.6$	$82.4^{a} \pm 3.3$	$111.8^{a} \pm 9.4$
N _{tot}								
Org. layer	$[mg g^{-1}]$	$7.8^{\rm D}(\pm 0.46)$	$8.8^{\mathrm{D}} \pm 0.36$	$14.1^{BC} \pm 0.56$	$13.1^{BC} \pm 0.61$	$11.9^{\circ} \pm 0.51$	$15.4^{AB} \pm 0.7$	$16.7^{A} \pm 0.57$
0–10 cm	[mg g ⁻¹]	$1.6^{\circ}(\pm 0.12)$	$1.4^{c} \pm 0.09$	$3.2^{b} \pm 0.36$	$3.1^{b} \pm 0.34$	$3.5^{\rm b}\pm0.18$	$4.7^{a} \pm 0.17$	$6.1^{a} \pm 0.48$
C/N								
Org. layer	$[g g^{-1}]$	$18.6^{AB} \pm 0.37$	$19.8^{A} \pm 0.42$	$17.9^{BC} \pm 0.21$	$19.3^{A} \pm 0.22$	$17.7^{BC} \pm 0.28$	$17.9^{BC} \pm 0.23$	$18.3^{ABC} \pm 0.14$
0–10 cm	$[g g^{-1}]$	$15.2^{d} \pm 0.51$	$19.8^{ab} \pm 0.75$	$16.2^{cd} \pm 0.51$	$19.9^{a} \pm 0.40$	$16.3 d \pm 0.30$	$17.5^{bc} \pm 0.16$	$18.3^{ab} \pm 0.23$
P _{tot}								
Org. layer	[mg g ⁻¹]	0.82 ± 0.04	0.96 ± 0.04	0.92 ± 0.04	0.92 ± 0.04	0.87 ± 0.04	0.90 ± 0.04	0.83 ± 0.05
P _{resin}								
0–10 cm	$[mg g^{-1}]$	$0.04^{a} \pm 0.003$	$0.01^{b} \pm 0.003$	$0.01^{b} \pm 0.003$	$0.01^{b} \pm 0.002$	$0.02^{b} \pm 0.002$	$0.01^{b} \pm 0.001$	$0.03^{a} \pm 0.003$

Information on soil texture was provided by the local forestry offices (Waldeckische Domanialverwaltung)

Geological epochs: l BU – lower Bunter (Triassic), l K – lower Keuper (Triassic), u D or m D—upper/middle Devonian shale P_{resin} – resin-exchangeable P

and after a full diurnal (photosynthetic) cycle of approx. 24 h, these trap solutions containing exudates were collected for analysis. These solutions were subsequently filtered through sterile glass fiber filters (GE Healthcare Life Sciences Whatman, Glass Microfibre Filters, Grade GF/F) and stored at -20 °C until further analysis. Control samples were taken from rootless cuvettes treated similarly. The samples were analyzed for their dissolved organic carbon using a total carbon analyzer (Shimadzu TOC-L CPH/CPN; Shimadzu Scientific Instruments, Duisburg, Germany). Taking fresh root biomass as a calculation basis, net mass-specific exudation rates (μ g C g⁻¹ h⁻¹) and annual C fluxes with exudation per

root mass or ground area (mg g⁻¹ yr⁻¹, g C m⁻² g⁻¹) were calculated, the latter from site-specific exudation-temperature relationships and temperature variation across the growing season, taking all days with average temperatures > 10 °C into account. This is a rough estimate, as it assumes a constant temperature dependence of exudation across the growing season. To estimate growing season length for the seven sites, we used gridded temperature data provided by the German Weather Service (DWD). Growing season length decreased with decreasing MAT from 170 to 125 d between 310 and 800 m a.s.l. (Table 1).

Fine root biomass and root morphology

In November 2018, 12 root samples were taken at random locations in each of the 30 m \times 30 m plots using a soil corer (6.6 cm in diameter) and the material separated into organic layer and mineral topsoil (0-10 cm) material. Samples were transported in a cooling box to the laboratory where they were kept at 4 °C and processed within four weeks. Only fine roots (diameter < 2 mm) of beech were considered in the analysis. All fine root segments were picked out by hand and sorted into live and dead fine root mass under a stereomicroscope (40×magnification). Root vitality was assessed by means of root color and structure of the root surface, root elasticity and turgescence, branching structure, and the degree of cohesion of cortex, periderm, and stele (for criteria, see Persson, 1978; Meier and Leuschner 2008). Standing FRB was expressed as profile total (organic layer and uppermost 10 cm of mineral soil; in g m^{-2}). Specific root length (SRL, m g⁻¹), specific root surface area (SRA, $\text{cm}^2 \text{g}^{-1}$), root tissue density (TD, mg cm⁻³), and root tip frequency (RTF, number of root tips per fine root mass; n g⁻¹) were determined for the root material using a flatbed scanner and the software WinRhizo (Régents Instruments, Quebec, Canada). Fine roots used for exudate collection were clipped off and the biomass of that root segment determined by drying (48 h, 78 °C) and weighing the sample.

Statistical analyses

All statistical analyses were conducted with SPSS software. The data was tested for fit to normal distribution using a Shapiro–Wilk test. Normally distributed data were tested for homogeneity of variances with a Levene test. Site differences between means of edaphic (total C and N content, C/N-ratio, pH (H₂O), pH (CaCl₂), total and plant-available P content) and root morphological variables (TD, SRA, SRL, and RTF), and exudation rates (net mass-specific exudation rate and annual C flux with exudation) were examined with one-way analysis of variance for parametric data and a Kruskal–Wallis test for non-parametric data. ANOVAs were followed by a Scheffé or a Dunnett T3 test, if homogeneity of variances was not fulfilled. Kruskal–Wallis tests were followed by pairwise comparisons to locate differences.

Pearson correlations were used for investigating the relation between root exudation rate and elevation, climatic and edaphic variables, and root morphological traits. If data were non-normally distributed, Spearman rank correlation analysis was employed. Correlations were tested for the variables long-term mean temperature and precipitation, average summer temperatures of 2014 and 2015, soil water content at the date of exudate sampling, and air and soil temperatures averaged over the seven days prior to sampling. The *p*-values were adjusted by the Benjamini–Hochberg procedure for multiple testing.

Multiple regression analyses with backward variable elimination were conducted to test for significant independent predictors of root exudation rates and the estimated annual C flux with exudation per m² ground area. As variation in soil chemical factors across sites was moderate to low due to the sampling design, and root morphology varied little, we ran the initial model with the abiotic factors site-specific daily temperature, soil moisture and elevation, and diameter at breast height and stem density as key characteristics of stand structure. At each elimination step, the variable showing the smallest contribution to the model was deleted until all variables remaining in the model produced significant F statistics. The p-values were calculated via the bootstrapping method because most of the data showed no fit to normal distribution. Variables were tested for multi-collinearity and were excluded when they were highly correlated and collinearity diagnostics (variance inflation factor and tolerance) were critical.

Results

Elevational change in climatic and edaphic conditions

During the two years of measurement, annual mean air temperature (weather station data)

decreased at a lapse rate of c. 0.57 °C per 100 m from 9.9 to 7.5 °C in 2014 (and from 9.3 to 6.9 °C in 2015) along the elevational gradient. Both years were on average by about 1.5 °C (2014) and 1.0 °C (2015) warmer than the long-term average of the 1981–2010 period (Tab. 1). Annual precipitation increased from 706 mm at 310 m to 1115 mm at 800 m in 2014, and from 571 to 1054 mm in 2015, corresponding to average increases by 83 and 99 mm per 100-m increase in elevation during the two study years.

The recording of climatic conditions in the stands directly before exudate sampling in summer 2015 gave smaller air temperature (T_a) and soil temperature (T_s) decreases with elevation during the sampling campaigns (T_a decrease by 0.37 and 0.53 °C 100 m⁻¹ and T_s decrease by 0.41 and 0.45 °C 100 m⁻¹ in August and September 2015, respectively; Fig. 1a). The amount of precipitation recorded one week prior to each sampling campaign showed a very slight decrease with elevation in 2014, with the lowest amount measured at the highest sites at 800 m (244 mm) and 690 m (260 mm) and the highest amount at the intermediate sites at 490 m (291 mm) and 560 m (281 mm). Precipitation increased with elevation during the August and September 2015 sampling campaigns (Fig. 1b). Gravimetric soil water content (SWC) of the organic layer plus uppermost mineral soil (0-10 cm) determined for the sampling in August and September 2015 roughly doubled between the lowest and highest sites from 15-25 wt. % at the low-elevation sites to 38-47 vol. % at the highelevation sites (Fig. 1c). In summer 2015, SWC increased along the gradient from 46.9% at 310 m to 67.0% at 800 m in the organic layer, and from 22.8% to 43.6% in the mineral topsoil (Table 2).

Among the most conspicuous changes observed in soil chemical and physical factors was the doubling in organic layer thickness from about 5.5 cm at 310 m to 10.5 cm at 800 m a.s.l., which was associated with an increase in organic layer mass from 8.0 kg m⁻² at 310 m to 15.0 kg m⁻² at 800 m (Table 2) and concomitant increases in organic layer C and N stocks. The organic carbon concentration (SOC) in the organic layer material increased from 143.7 mg g⁻¹ at 310 m to 304.4 mg g⁻¹ at 800 m, and that of the mineral topsoil from 26 mg g⁻¹ at 310 m to 111.83 mg g⁻¹



Fig. 1 Soil and air temperature (**a**), precipitation (**b**) and soil water content (SWC; wt.%) (**c**) measured in seven forests at two sampling dates in summer/autumn 2015 and one sampling date in summer 2014 (precipitation only) (means \pm SE). Temperatures are means of seven-day periods prior to exudate sampling of each 5 sensors per stand, precipitation data are interpolated from gridded weather station data of the German Weather Service (DWD) corrected for elevation, and SWC data are gravimetric samples taken in the organic layer and 0–10 cm mineral soil at the date of sampling (n=5)

at 800 m. In parallel, total N concentration increased in the organic layer from 7.8 mg g⁻¹ at 310 m to 16.7 mg g⁻¹ at 800 m, and in the mineral topsoil from 1.64 mg g⁻¹ at 310 m to 6.05 mg g⁻¹ at 800 m. Soil C/N ratio and P content showed no elevational trends in both layers (Table 2).

Elevational change in fine root biomass and root morphology

FRB in the organic layer increased with elevation largely in parallel with the increasing forest floor depth from 2.2 g m⁻² at 310 m to 24.0 g m⁻² at 690 m (significant relation; r=0.51, p < 0.01), revealing a positive relation to MAP and a negative one to MAT (Table 3). FRB in the mineral topsoil (0–10 cm) varied between 9.3 and 48.9 g m⁻² without a clear elevational trend (Fig. 2). Both FRB components were negatively related to the P concentrations in the organic layer (P_{tot}) and mineral soil (P_{resin}) but were unrelated to soil C/N ratio. None of the examined root morphological parameters (SRL, SRA, TD, RTF) changed significantly with elevation, nor were influenced by climatic or soil chemical variables (Table 3) (Fig. 3).

Elevational change in exudation rates and dependence of exudation on climatic and stand structural properties

Average mass-specific root exudation rate across the seven sites was $18.34 \pm 6 \ \mu g \ g^{-1} \ h^{-1}$ (Fig. 4a) with site means (averaged over all sampling dates) ranging from 12.2 μ g g⁻¹ h⁻¹ at 690 m to 21.6 μ g g⁻¹ h⁻¹ at 380 m a.s.l. (peak rates > 37 μ g g⁻¹ h⁻¹). While lowest specific exudation rates were measured at the highest sites 690 and 800 m (12.2—15.0 μ g g⁻¹ h⁻¹), variation among sites was large and peak rates were recorded at mid elevation (490 m: 22.7 μ g g⁻¹ h⁻¹) and not at the lowest sites (Fig. 4a). Yet, the negative relation between mean exudation rate and elevation was highly significant (r=-0.31, p < 0.01; Tab. 3). C flux with exudation per ground area, calculated by multiplying specific exudation with FRB in the organic layer, was significantly higher above 400 m a.s.l. than at lower elevation (Fig. 4b).

Estimated annual cumulative C fluxes with exudation per root mass, derived from site-specific exudation-temperature relationships and recorded temperature variation across the growing season (all days with means > 10 °C), decreased significantly with elevation along the transect, from 80–85 mg g⁻¹ yr⁻¹at 310 m to 40–45 mg g⁻¹ yr⁻¹ at 800 m a.s.l. (mean of the seven sites: 66 ± 29 mg g⁻¹ yr⁻¹; Fig. 4c). Estimating annual C flux per ground area by multiplying the cumulative, mass-specific C flux with the FRB total of the organic layer and mineral topsoil yields a different elevational pattern. Highest standlevel C fluxes were calculated for 380 m (ca. 4.6 g C $m^{-2} yr^{-1}$) and 600 m (ca. 3.2 g $m^{-2} yr^{-1}$) with highest FRB, and lowest for 310 m (ca. 0.8 g $m^{-2} yr^{-1}$) and also 490 m (ca. 1.6 g $m^{-2} yr^{-1}$), but without a dependence on elevation (Fig. 4d). Averaged over all sites, we calculated a mean growing season C flux of 2.2 ± 0.2 g m^{-2} .

The environmental factors with strongest influence on mass-specific root exudation rate were the site-specific daily air and soil temperatures with highly significant positive effects (r = 0.66, p < 0.01; Table 3a), revealing a linear increase of exudation rate at a slope of 1.97 $\mu g g^{-1} h^{-1}$ per 1 °C temperature increase (Fig. 5a; Fig. S1 in the Supplement). In contrast, the relationships to MAT and MST and mean temperature of the measuring year were less tight (r = 0.29 - 0.31, p < 0.01 or n.s.). Close relations were also found for soil moisture (SWC) and MAP (both negative), and the N content of the upper soil (negative) and soil pH (CaCl₂) (positive), while soil P content and C/N ratio were not influential (Table 3a). In contrast to actual exudation rate, the annual stand-level C flux depended only on the Presin concentration of the topsoil (negative relation; r = 0.25; p < 0.05) and specific root area (positive relation; r = 0.25; p < 0.05), but not on any climatic factor. Annual fluxes consequently were related neither to actual (average summer) temperature nor to long-term mean summer temperature (MST) (Fig. 5c and d).

While FRB decreased with increasing stem density (r=0.36), specific root exudation rate did increase (r=0.32); Table 3b and Fig. S2 in the Supplement).

Multiple regression analyses on climatic and stand structural drivers of root exudation

Multiple regression analyses on the influence of various climatic and stand structural factors on the exudation rate of beech indicate a dominant positive effect of the site-specific daily air temperature (F=69.7; p < 0.001), followed by a somewhat weaker negative effect of diameter at breast height (DBH; F=39.3; p < 0.001). The model explained 41% of the variance in exudation rate (Table 4). Annual C flux per ground area was influenced only by DBH, but not by temperature or any other climatic variable; the explained variance was only 14%.

Table 3 (a) Results of Spearman rank correlation analyses
on the dependence of fine root biomass (FRB), root morpho-
logical parameters (specific root length-SRL, specific root
surface area-SRA, root tissue density-TD and root tip fre-
quency-RTF), mass-specific root exudation rate, and extrapo-
lated annual C flux with exudation on elevation, mean annual
precipitation (MAP), mean annual temperature (MAT), multi-
annual summer temperature (MST), mean soil water content
(SWC) in summer, precipitation immediately before sam-
pling, and air and soil temperature 7 days prior to sampling,
and various soil chemical parameters in the organic layer,
mineral topsoil (0-10 cm) and of the pooled organic layer and

mineral topsoil (n=5 measurements per layer and site, n=2 sampling dates in 2015). (b) Results of Spearman rank correlation analyses on the dependence of fine root biomass (FRB), root morphological parameters (SRL, SRA, TD and RTF), and mass-specific root exudation rate and extrapolated annual C flux with exudation on stem density, SRL, SRA, TD and RTF. Given are the Spearman correlation coefficients and the significance of the relationship (*: p < 0.05, **: p < 0.01). Positive correlations are indicated by positive r values, negative ones by negative r values. *p*-values were adjusted by the Benjamini–Hochberg procedure for multiple comparison. Correlation coefficients > 0.5 are printed in bold

		FRB org.lay	FRB 0-10 cm	SRL	SRA	TD	RTF	Root exudation rate	Annual C Flux
		(g m ⁻²)	$(g m^{-2})$	(cm g^{-1})	$(\mathrm{cm}^2~\mathrm{g}^{-1})$	(mg cm ³)	$(n \ g^{-1})$	$(\mu g \; g^{-1} \; h^{-1})$	$(g m^{-2} yr^{-1})$
a)									
Elevation	(m a. s. l.)	0.510**	0.093	-0.032	0.032	-0.013	0.082	-0.305**	-0.070
MAP (1981– 2010)	(mm)	0.510**	0.093	-0.032	0.032	-0.013	0.082	-0.305**	-0.075
MAT (1981– 2010)	(°C)	-0.466**	0.104	0.038	-0.044	0.009	-0.069	0.305**	0.116
MST (1981– 2010)	(°C)			0.097	0.076	-0.005	0.003	0.291	0.116
Ann. precip. 2014, 2015	(mm)	-	-	-0.024	0.009	0.113	0.087	-0.288**	-0.063
Ann. temp. 2014, 2015	(°C)	-	-	0.049	-0.041	0.068	-0.039	0.313**	0.116
Mean SWC (topsoil)	(wt%)	-	-	-0.014	0.077	-0.043	0.113	-0.495**	-0.187
Actual precipita- tion	(mm)	-	-	-0.022	-0.016	0.182	0.051	-0.169	-0.211
Daily air tempera- ture	(°C)	-	-	0.092	0.066	-0.116	-0.056	0.669**	0.068
Daily soil tem- perature	(°C)	-	-	0.090	0.057	-0.097	-0.060	0.656**	0.185
C _{tot}									
Org. layer	$[mg g^{-1}]$	0.496**	-	-	-	-	-	-	-
0–10 cm	$[mg g^{-1}]$	-	-0.023	-	-	-	-	-	-
Upper soil	$[mg g^{-1}]$	-	-	-0.042	0.038	-0.024	0.058	-0.303**	-
Upper soil	[g m ⁻²]	-	-	-	-	-	-	-	-0.059
N _{tot}									
Org. layer	$[mg g^{-1}]$	0.496**	-	-	-	-	-	-	-
0–10 cm	$[mg g^{-1}]$	-	-0.042	-	-	-	-	-	-
Upper soil	$[mg g^{-1}]$	-	-	-0.044	0.061	-0.055	0.058	-0.320**	-
Upper soil	[g m ⁻²]	-	-	-	-	-	-	-	-0.179
C:N									
Org. layer	$[g g^{-1}]$	0.037	-	-	-	-	-	-	-
0–10 cm	$[g g^{-1}]$	-	0.315	-	-	-	-	-	-
Upper soil	$[g g^{-1}]$	-	-	0.081	0.060	0.04	0.014	0.080	0.075
pH H ₂ O									
Org. layer		-0.373**	-	-	-	-	-	-	-
0–10 cm		-	0.103	-	-	-	-	-	-

Table 3 (continued)

			FRB org lay	FRB 0-10 cm	SRL	SRA	TD	RTF	Root exudation rate	Annual C Flux
			$(g m^{-2})$	$(g m^{-2})$	(cm g^{-1})	$(cm^2 g^{-1})$	(mg cm ³)	(n g ⁻¹)	$(\mu g \ g^{-1} \ h^{-1})$	$(g m^{-2} yr^{-1})$
Upper soi	1		-	-	0.015	-0.078	0.023	-0.068	0.198	0.116
pH (CaCl ₂)										
Org. layer			-0.432**	-	-	-	-	-	-	-
0–10 cm			-	0.037	-	-	-	-	-	-
Upper soi	1		-	-	0.013	-0.052	0.006	-0.075	0.244**	0.129
P _{tot}										
Org. layer		$[mg g^{-1}]$	-0.345**	-	0.147	0.024	0.125	0.075	0.204	-
Org. layer		[g m ⁻²]	-	-	-	-	-	-	-	0.045
P _{resin}										
0–10 cm		$[mg g^{-1}]$	-	-0.326**	0.061	0.015	0.145	0.105	0.020	-
0–10 cm		[g m ⁻²]		-	-	-	-	-	-	-0.246*
b)										
stem dens	ity	$(n ha^{-1})$	-0.360**	0.073	0.081	0.014	0.029	-0.033	0.323**	0.135
SRL		(cm g^{-1})	-0.020	0.046	-	-	-	-	-	-
SRA		$(cm^2 g^{-1})$	0.068	0.049	-	-	-	-	-	-
TD		(mg cm ³)	-0.079	-0.096	-	-	-	-	-	-
RTF		$(n g^{-1})$	0.084	0.017	-	-	-	-	-	-

Fig. 2 Fine root biomass of beech in the organic layer (grey bars) and the mineral topsoil (0–10 cm) of the seven beech forests in November 2018 (means \pm SE of 12 samples per layer). Different letters indicate significantly different means of organic layer samples (capital letters) and mineral soil samples (small letters)



Discussion

The measured mass-specific root exudation rates of the mature beech trees in our study (site means of 12–22 μ g C g⁻¹ h⁻¹) were similar to rates reported

in other studies for beech in moist soil (10–23 μ g C g⁻¹ h⁻¹, Liese et al. 2018; 16–65 μ g C g⁻¹ h⁻¹, Meier et al. 2020) and corresponded also to values found in other tree species in the temperate zone (e.g., black locust: 10–22 μ g C g⁻¹ h⁻¹, Uselman et al. 2000;



Fig. 3 Specific root length, specific root area, root tissue density and root tip frequency of the fine root biomass samples used for exudation collection in the seven forests. Shown are means and standard errors of n=9 samples per site and sampling date with averaging over the three sampling dates

loblolly pine: $12-26 \ \mu g \ C \ g^{-1} \ h^{-1}$, Meier et al. 2013). Our simple and multiple regression analyses indicate that the thermal conditions during and 7 days prior to

sampling are an important factor controlling exudation of beech across the studied elevational gradient. Exudation per root mass increased by about 2 μ g C g⁻¹ h⁻¹ per 1 °C temperature increase. In two Mediterranean tree species, Jakoby et al. (2020) found a linear increase in exudation rate per root surface area of about 0.15 μ g C cm⁻² d⁻¹ per 1 °C increase, and soil temperature together with soil moisture explained exudation dynamics best in one of the species. Uselman et al. (2000) found a 70% higher root exudation in *Robinia pseudoacacia* seedlings when temperature was increased from 26 to 30 °C.

Several explanations of a positive temperature effect on exudation are possible, which relate to the 'push' and 'pull' hypotheses of the control of C acquisition by roots (Farrar et al. 2003). In the humid climate of our study region, higher summer temperatures are usually related to higher insolation and thus greater carbon assimilation, as beech photosynthesis is primarily limited by radiation at montane elevation (Schulze 1970). Consequently, under warmer conditions, more C should in principle be available for allocation to roots, which might enhance root exudation due to higher levels of non-structural carbohydrates in roots (Prescott et al. 2020) and a steeper concentration gradient of soluble organic compounds between root cells and the soil (Jones et al. 2004). The photosynthetic capacity of beech at montane elevation in Central Europe reveals a pronounced seasonality with a peak in the warmest months (Schulze 1970), which might imprint on the seasonality of exudation, as photosynthetic capacity and root exudation have been found to be closely related (Sun et al. 2017). Another possible explanation focuses on the role of root exudation for the metabolic activity of rhizosphere biota and its stimulating effect on nutrient supply for root uptake (Jones et al. 2004). Warmer weather likely increases the plant demand for nitrogen and other nutrients as photosynthetic capacity is ramped up, which might trigger roots to stimulate soil microbial activity through active secretion of labile C as an energy source for microbes (Pausch and Kuzyakov 2017). Such a mechanism would fit to the 'pull' hypothesis, as exudation then were primarily controlled by factors other than C supply from the canopy (Karst et al. 2017). It is possible that both 'push' and 'pull' mechanisms are underlying the observed increase in exudation with rising temperature.





Fig. 4 Means (and SE) of (**a**) measured average mass-specific root exudation rate, (**b**) daily exudation for the fine root mass in the organic layer expressed per m^2 ground area, (**c**) extrapolated annual mass-specific carbon flux with exudation, and (**d**) estimated annual C flux with exudation per m^2 ground area in the topsoil of the seven beech forests along the elevation gradient (averaged over n=9 samples taken per date and site and n=3 sampling dates in 2014 and 2015). Annual exudation (**c** and **d**) was estimated from the site-specific exudation-temperature relationship, temperature variation across the vegetation

The fact that exudation was responsive to increases in the actual air and soil temperature at the site but showed no significant relation to mean growing season temperature (MST), i.e. the long-term average thermal conditions of the site, may suggest thermal acclimation of exudation to the mean temperature at the sites. Since other processes of plant carbon turnover such as respiration and photosynthesis are subject to marked thermal acclimation (Atkin and Tjoelker 2003; Hikosaka et al. 2006) and exudation depends on photosynthetic C gain, we would indeed expect that beech trees growing at lower elevation in a warmer climate down-regulate their exudation rate to a certain degree, at least that fraction of exudates that is

period, and the specific length of the vegetation period (number of days with mean temperature ≥ 10 °C) at the sites. As we calculated annual totals with pooled data, no SD can be given in figures c) and d). The exudation flux per ground area was extrapolated using the fine root biomass data from the organic layer (b) or the organic layer and the mineral topsoil (0–10 cm) (d). Different capital letters denote significantly different means at p < 0.05 with *p*-values adjusted by the Benjamini–Hochberg procedure for multiple comparisons (Kruskal–Wallis test)

controlled by the plant. Mean exudation rates would then become more similar across the elevational gradient. More data from field studies along temperature gradients are needed to test this hypothesis.

Since our model considers only several climatic, soil hydrologic and stand structural factors, while ignoring other likely influencing factors (such as photosynthetic activity and mycorrhizal infection) and covering other drivers only marginally due to limited parameter variation in our sample (notably soil N and P availability), we cannot assess the relative importance of temperature as a determinant of root exudation in beech. However, the positive relation of exudation to temperature is remarkably strong with a



Fig. 5 Relationships between site-specific daily air temperature (means of the seven days prior to exudate measurement; left panels) or long-term mean summer air temperature (right panels) and (a) and (d) mass-specific root exudation rates, (b) and (e) cumulated annual C exudation per root mass, and (c) and (f) cumulated annual C exudation per ground area in the seven beech forests along the elevation gradient in summer

Table 4 Multiple regression analyses with backward variable elimination on the effects of elevation, site-specific daily air temperature, soil water content (SWC) and elevation, and diameter at breast height and stem density on root exudation

2015 (averaged over n=9 samples per date and site; only 2015 data considered). Temperature data were measured with I-button loggers at the sampling sites. Annual rates take the variable length of the vegetation period (no. of days ≥ 10 °C) at the sites into account. Different colour of symbols indicates elevation of sites

rates and the estimated annual carbon flux per m^2 ground area in the topsoil of the seven beech forests along the elevational gradient. The ± signs indicate positive/negative correlation

Y	Model			Predictor	F	Р
	$\overline{R^2}$	Р				
Exudation rate	0.41	< 0.001	<0.001 + Site-specific daily air temperature	Site-specific daily air temperature	69.7	< 0.001
			-	DBH	39.3	< 0.001
Annual C flux per ground area	0.14	0.001	-	DBH	13.1	0.001

correlation coefficient of 0.66, suggesting an important role for summer temperatures and the duration of warm summer periods for exudation in beech. Soil moisture (SWC) is another abiotic factor with a presumably large effect on exudation. Low to moderate drought seems to increase exudation

(Jakoby et al. 2020; Liese et al. 2018; Preece et al. 2018), apparently as a stress response similar to the exudation increase observed upon soil cooling to 4 °C (Karst et al. 2017). The seven beech forests in our study represent a precipitation and soil moisture gradient with annual precipitation in the study year 2015 decreasing to nearly a half (1050 - 570 mm) and mean soil moisture in summer to less than a half (ca. 45 to ca. 20 wt.%) from 800 to 310 m elevation. In fact, the highest exudation rates were measured at the driest sites and the negative relationship to mean soil water content was highly significant, which may offer an alternative explanation of the exudation pattern along the slope, apart from a positive temperature effect. However, the correlation of exudation to soil moisture was less tight than to temperature, and SWC was not included in the multiple regression model. Moreover, mid and late-summer moisture contents in the topsoil of 15-25 wt.% likely have exposed the trees to only mild drought, which may question that exudation was stimulated by drought in our sample. It thus remains unclear, how important water availability is for the explanation of the observed exudation patterns.

Soil nutrient availability is another abiotic factor with a possible influence on the exudation rate. For example, P deficiency results in enhanced root secretion of phenolic compounds in certain species (Neumann and Römheld 1999), and exudation of beech is higher at acidic, N-poor sites than in more fertile soils (Meier et al. 2020), which may trigger increases in decomposition rate and thus N availability through a rhizosphere priming effect (Cheng et al. 2014; Jones et al. 2004; Phillips et al. 2011). The negative correlation between topsoil N content and exudation rate in our sample suggests a stimulating effect of N deficiency on exudation. There are, however, also reports of no influence of soil nitrogen deficiency on exudation, or stimulation by N addition (Uselman et al. 2000; Yin et al. 2013b). Along the studied elevational gradient, organic layer mass on the forest floor doubled from 310 to 800 m a.s.l. with the consequence that topsoil C and N pools increased, whereas C/N ratio, P_{tot} and P_{resin} content as well as soil pH did not change, suggesting that plant-availability of N and P varied only little. Gradients with larger variation in N and P availability are needed to study the relative importance of nutrient availability on the hand, and temperature on the other, on exudation in beech.

We found only minor alterations in beech fine root morphology from 310 to 800 m a.s.l.; yet, the structure of the fine root system changed markedly. FRB in the organic layer increased roughly tenfold in parallel with the increasing depth of this layer, while FRB in the mineral topsoil varied without a clear elevational trend. We interpret this pronounced shift of FRB to the surface layer as a consequence of decreasing litter decomposition rates toward higher elevations, prompting the beech trees to concentrate their fine root mass in the organic surface layers with highest mineralization rates. This change in root distribution patterns was associated with apparent change in exudation per unit soil volume: While mass-specific exudation rate was somewhat lower, FRB was much larger in the topsoil of the high-elevation beech forests with the consequence that daily exudation rates per topsoil volume were significantly higher above 400 m elevation where the climate is cooler. Higher root densities in colder environments have been observed in northern as well as high-elevation forests (Helmisaari et al. 2007; Kubisch et al. 2016; Moser et al. 2011). They might serve two purposes, to increase the absorbing root surface area under conditions of reduced nutrient supply and to enhance the stimulation of microbial activity under low temperatures. Lower mineralization rates in cooler soil should increase the necessity for the trees to conduct rhizosphere priming by stimulating soil microbial activity through the provision of labile C as an easily accessible energy source.

Conclusions

Our study of root exudation of mature beech trees in seven forests along an elevation (and associated temperature) gradient provides evidence that mass-specific exudation increases with temperature at the time of measurement, whereas the dependence on the site's average summer temperatures was weak. Although we fully recognize the limitations of our data set with an only short MST gradient and only three measuring campaigns at the seven sites, our results are convincing with respect to the prominent positive temperature effect on exudation, which was more important than effects of soil moisture, precipitation or stand structure. With data on FRB at the sites and information on the length of the growing season, we were able to compare the seven forests with respect to calculated exudation per ground area and to give rough estimates of cumulative exudation per growing season. Both a higher mass-specific exudation rate and a longer growing season length contribute to a generally higher cumulative exudation at the warmer sites. Future research should investigate whether higher exudation and thus more intense rhizosphere priming is a factor that contributes to the generally higher availability of N, P, and other nutrients in soil when temperature increases. Studies on temperaturedependent changes in the composition of exuded substances are another promising field of future study. In combination with earlier studies in European beech forests across edaphic and climatic gradients (Meier et al. 2020; Tückmantel et al. 2017), this study deepens our understanding of environmental controls of root exudation in this model tree species.

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Code availability Not applicable

Declarations

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